

Soil organic matter in fire-affected pastures and in an Araucaria forest in South-Brazilian Leptosols

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Abstract – The objective of this work was to evaluate the distribution pattern and composition of soil organic matter (SOM) and its physical pools of Leptosols periodically affected by fire over the last 100 years in South Brazil. Soil samples at 0–5, 5–10, and 10–15 cm depths were collected from the following environments: native pasture without burning in the last year and grazed with 0.5 livestock per hectare per year (1NB); native pasture without burning in the last 23 years and grazed with 2.0 livestock per hectare per year (23NB); and an Araucaria forest (AF). Physical fractionation was performed with the 0–5 and 5–10 cm soil layers. Soil C and N stocks were determined in the three depths and in the physical pools, and organic matter was characterized by infrared spectroscopy and thermogravimetry. The largest C stocks in all depths and physical pools were found under the AF. The 23NB environment showed the lowest soil C and N stocks at the 5–15 cm depth, which was related to the end of burning and to the higher grazing intensity. The SOM of the occluded light fraction showed a greater chemical recalcitrance in 1NB than in 23NB. Annual pasture burning does not affect soil C stocks up to 15 cm of depth.

Index terms: physical pools, carbon stock, chemical composition, chemical recalcitrance, thermo-decomposition.

Matéria orgânica do solo em pastagens afetadas pelo fogo e em mata de Araucária em Neossolos sul-brasileiros

Resumo – O objetivo deste trabalho foi avaliar o padrão de distribuição e a composição da matéria orgânica do solo (MOS) e de seus compartimentos físicos em um Neossolo periodicamente afetado pelo fogo nos últimos 100 anos, no Sul do Brasil. Amostras de solo foram coletadas nas profundidades de 0–5, 5–10 e 10–15 cm, nos seguintes ambientes: pastagem nativa sem queima há um ano e com lotação de 0,5 animal por hectare por ano (1NB); pastagem nativa sem queima há 23 anos e com lotação de 2,0 animais por hectare por ano (23NB); e mata de Araucária (AF). Fracionamento físico foi realizado com amostras das camadas de 0–5 e 5–10 cm. Os estoques de C e N do solo foram determinados nas três profundidades e nas frações físicas, e a matéria orgânica foi caracterizada por espectroscopia de infravermelho e termogravimetria. O maior estoque de C em todas as profundidades e as frações físicas foi verificado na AF. O ambiente 23NB apresentou o menor estoque de C e N do solo na profundidade de 5–15 cm, associado ao cessar fogo e à maior densidade de pastejo. A MOS da fração leve oclusa apresentou maior recalcitrância química em 1NB do que em 23NB. A queima anual da pastagem não afeta o estoque de C até 15 cm de profundidade.

Termos para indexação: frações físicas, estoque de carbono, composição química, recalcitrância química, termo-decomposição.

Introduction

Native pastures occupy approximately 23% of the Rio Grande do Sul State area, in Brazil, and provide forage for cattle and sheep breeding (Hasenack et al., 2007). Pasture management systems are mainly based on the intensity and frequency of grazing, grassland cutting,

soil liming and fertilization, and vegetation burning. Although strongly questioned by environmentalists and restricted by legislation, pasture burning after winter is still employed in some highland regions of the state of Rio Grande do Sul. The purpose of rapid burning is to stimulate the regrowth of pasture and to increase the quality of the produced forage (Damé et al., 1997).

Vegetation burning may promote an increase or decrease of soil organic matter (SOM) content depending on fire intensity and duration, environmental conditions, management systems, and other factors (Santín et al., 2008). Pasture burning is frequently associated with animal overstocking (Bertol et al., 1998), and, depending on grazing intensity, the combined management systems may promote the reduction of the soil carbon concentration (Abril et al., 2005). In an Oxisol under pasture in southern Brazil, the cessation of burning for eight years restored soil C stocks to the same levels found under native pasture unaffected by fire for the last 41 years (Santana, 2010).

In addition to the effect on SOM stocks, vegetation burning may also change the chemical composition of SOM and produce black carbon. In a Haplic Luvisol under pasture, black carbon was observed within microaggregates (Brodowski et al., 2006). According to these authors, the enrichment of black carbon in these structures contributed to the formation and stabilization of microaggregates. Similarly, in Andosols, the presence of black carbon was identified in particulate and occluded fractions, whose chemical composition was highly aromatic (Golchin et al., 1997). One year after burning, black carbon-derived compounds were identified in the B horizon of Spanish soils, as confirmed by an increase in aromatic C content, indicating that some pyrogenic compounds may be transported into deeper soil horizons (Knicker et al., 2006).

The SOM distribution in physical pools may provide relevant information about the impact of the pasture management system on its dynamics. The light fractions (free, occluded, and particulate $>53 \mu\text{m}$) are usually highly sensitive to changes in soil management. The SOM free light fraction is stabilized mainly by its chemical recalcitrance, whereas, in the occluded light fraction, both physical protection and organo-mineral association mechanisms act in SOM stabilization (Christensen, 2001). Moreover, the contents and quality of the physical pools are usually useful tools to evaluate changes in SOM dynamics caused by the different management systems (Paul et al., 2008).

According to a previous study on pasture management in a Leptosol, vegetation burning after winter promoted an increase in the soil organic carbon (SOC) concentration and a higher proportion of biochemically labile structures, such as carbohydrates,

in comparison to pastures that were not exposed to fire (Potes et al., 2010).

The objective of this work was to evaluate the distribution pattern and composition of SOM and its physical pools of Leptosols periodically affected by fire over the last 100 years in South Brazil.

Materials and Methods

The experimental areas are located in São José dos Ausentes, Rio Grande do Sul State, Brazil ($28^{\circ}44'52''\text{S}$ and $50^{\circ}03'57''\text{W}$, at approximately 1,200 m above sea level). The climate is mesothermal, with an annual mean temperature of 14.4°C and annual rainfall of 2,468 mm (Streck et al., 2008). The soil is classified as a Neossolo (Santos et al., 2006) or Leptosol (Soil Taxonomy) and has a clayey silty texture to a 5 cm depth and a clayey texture at the 5–15 cm layer. The native vegetation represents a transition zone between grassland and *Araucaria* sp. forest (Heringer & Jacques, 2002). The pasture areas have been subjected to annual burning immediately after winter for at least the past 100 years, and no liming or fertilization has ever been performed. The burning of the dried vegetation occurs rapidly due to the low humidity and temperature of the soil and vegetation (lower than 10°C) at the end of winter (August and September). Both the pasture and the *Araucaria* forest areas have been under native vegetation since the colonization of the region in the 18th century. The management systems consisted of: native pasture without burning in the last year and grazed with 0.5 livestock per hectare per year (1NB) ($28^{\circ}36'4.4''\text{S}$ and $49^{\circ}58'35.3''\text{W}$); native pasture without burning in the last 23 years and grazed with 2.0 livestock per hectare per year (23NB) ($28^{\circ}35'26.9''\text{S}$ and $49^{\circ}52'30.5''\text{W}$); and *Araucaria* forest (AF) ($28^{\circ}36'1.9''\text{S}$ and $49^{\circ}58'36.4''\text{W}$). The original condition under *Araucaria* forest was used as a reference.

Soil sampling was performed at the end of winter just before fire-setting. In each management system, three pits of $1.0 \times 1.0 \times 0.20$ m depth were opened, and undisturbed soil samples were collected in three soil layers: 0–5, 5–10, and 10–15 cm. The sampled soils were passed through a 9.51 mm sieve. The average particle size distribution (determined by the pipette method) consists of: 430 g kg^{-1} of clay, 427 g kg^{-1} of silt, and 143 g kg^{-1} of sand in the 0–5 cm layer; 467 g kg^{-1}

of clay, 393 g kg⁻¹ of silt, and 140 g kg⁻¹ of sand in the 5–10 cm layer; and 503 g kg⁻¹ of clay, 353 g kg⁻¹ of silt, and 144 g kg⁻¹ of sand in the 10–15 cm layer.

The total soil C and N contents were determined by dry combustion (975°C) in each field replicate using a Perkin Elmer 2400 (PerkinElmer, Waltham, MA, USA). The C content was assigned exclusively to SOM, as all soil samples were carbonate free. The soil organic carbon (C_{soil}) and nitrogen (N_{soil}) stocks (Mg ha⁻¹) were quantified by correcting the concentration values using the 23NB environment as the baseline reference. This procedure normalizes the different effects the management systems have on soil bulk density from undisturbed soil cores (6.5 cm in diameter and 5.0 cm long). For this purpose, the following equation was used (Gatto et al., 2010): C_{soil} or $N_{soil} = (E_c \times \rho_b \times T) / 10$, in which C_{soil} or N_{soil} is the total soil organic carbon or total soil nitrogen stocks (Mg ha⁻¹), respectively; E_c is the element concentration (g kg⁻¹); ρ_b is the field bulk density (g cm⁻³); and T is the thickness of the soil layer (cm).

The density fractionation of the SOM was performed according to a method adapted from Conceição et al. (2008), with <9.51 mm soil aggregates of the 0–5 and 5–10 cm layers. The adaptation was necessary for the effective separation of the free light fraction (FLF) due to the resistance of the aggregates to absorbing the sodium polytungstate (SPT) solution. This behavior was caused by the high SOM content, which hindered the SPT absorption due to its hydrophobic character and to the blocking of pores. Consequently, air bubbles were trapped inside the aggregates, preventing them from being disrupted and from decanting. To avoid this problem, approximately 10 g of soil were wetted by capillarity prior to the density fractionation by laying the sample on a filter paper embedded in distilled water for 12 hours. The amount of absorbed water was gravimetrically determined, and, in the preparation of the SPT solution, this amount was considered to have a final density of 2.0 g cm⁻³ during the physical fractionation.

Approximately 80 mL of the SPT solution with a density of 2.0 g cm⁻³ was added to the wetted samples in a centrifuge tube. The tube was closed with a rubber stopper and gently inverted five times to release the FLF located between aggregates. The suspension was centrifuged at 2,000 g for 90 min, and the supernatant containing the FLF was filtered under vacuum in a

0.45 µm glass fiber filter (Whatman GF/A, Whatman, Inc., Clifton, NJ, USA), having been previously weighed. The material on the filter was washed with distilled water, followed by a 0.01 mol L⁻¹ CaCl₂ solution to remove the SPT solution, and finally by distilled water (200 mL) to remove the remaining CaCl₂. The filtered SPT solution was returned to the tube containing the aggregate pellet and the suspension was sonicated (Vibracell VC 750, Vibracell Comércio de Eletrônicos, São Paulo, SP, Brazil) at 400 J mL⁻¹ to break up the aggregates and, consequently, release the occluded light fraction (OLF).

After soil dispersion, the suspension was centrifuged and filtered following the procedure performed with the FLF. The residual soil (heavy fraction, HF) remaining in the tube after the separation of the light fractions was washed with distilled water (200 mL) and with a 0.01 mol L⁻¹ CaCl₂ solution. The separated fractions were oven-dried at 50°C and then weighed and ground. The C and N contents of the light fractions (FLF and OLF) were determined following the procedure used for all soil depths. The complete removal of SPT from the HF is not always possible; therefore, this fraction may remain contaminated by this reagent. The contents of C and N contained in the HF were determined by the difference between the element content in the soil and the respective content in the two light fractions. After discounting the weight of the filter, the C and N stocks (Mg ha⁻¹) of the physical fractions were calculated, a procedure that was also applied to the whole soil.

The physical fractions were treated with a 10% (v v⁻¹) hydrofluoric acid solution (Gonçalves et al., 2003), resulting in a concentrated SOM sample (FLF_{HF}, OLF_{HF}, and HF_{HF}). The organic fraction of the FLF and OLF was removed by treatment with hydrogen peroxide (H₂O₂) 30% (v v⁻¹) (Silva et al., 2008) to isolate the mineral fraction (FLF_{H₂O₂} and OLF_{H₂O₂}).

The untreated light fractions (FLF and OLF), the SOM concentrated samples (FLF_{HF}, OLF_{HF}, and HF_{HF}), and the peroxide-treated FLF and OLF (FLF_{H₂O₂} and OLF_{H₂O₂}) were analyzed by fourier transform infrared (FTIR) spectroscopy (Shimadzu FTIR 8300, Shimadzu, Kyoto, Japan) in KBr pellets (1.0 mg sample:100 mg KBr) using 32 scans and a resolution of 4.0 cm⁻¹ within the range of 4,000 to 500 cm⁻¹. The absorption band attributions were made according to Tan (1996).

The thermo-decomposition curves of the physical fractions (0–5 and 5–10 cm) were obtained using

a thermo-gravimetric analyzer (TGA-Q50, TA Instruments, New Castle, DE, USA) with 5 to 10 mg samples over static air. The initial weight was stabilized at 40°C, and a heating curve was obtained with a heating rate of 10°C min⁻¹ up to 105°C, with a holding time of 10 min, followed by heating at 5°C min⁻¹ up to 700°C. The FLF from the 0–5 cm layer was analyzed in the untreated form, since its yield after hydrofluoric acid treatment was small. To isolate the weight loss of the FLF in the temperature interval from 105 to 600°C due to the inorganic matrix, a TGA analysis was performed with the sample FLF_{H2O2}. The observed weight loss was then subtracted from that of the untreated FLF in the same temperature interval. The other OLF_{HF} and HF_{HF} samples were analyzed in the 0–5 and 5–10 cm layers. The thermo-decomposition curves were divided into four regions selected according to the sample decomposition behavior during heating: the first region was from 40 to 105°C; the second, from 105 to 350°C (1NB and AF) or 105 to 425°C (23NB); the third, from 350 to 600°C (1NB and AF) or 425 to 600°C (23NB); and the fourth region was from 600 to 700°C. The quotient of the weight loss between that of the third and the second region was calculated and defined as a thermogravimetric index (TGI) (Benites et al., 2005).

The results of the soil C and N stocks and density fractionation were evaluated by descriptive data analysis based on the mean standard error.

Results and Discussion

The SOC stocks varied between 26.2 and 52.0 Mg ha⁻¹ in the studied layers (Table 1). In the 1NB system, the SOC stocks were relatively uniform up to a depth of 15 cm. In contrast, the SOC stocks in the 23NB system tended to decrease with depth, from 31 to 26 Mg ha⁻¹. Furthermore, the SOC stock values in the 5–10 and 10–15 cm layers in 23NB were lower than those found under 1NB. This behavior can be related to the more intensive grazing in 23NB, which reduces the input of vegetal residue to the soil, in comparison to 1NB. In an Argentinean Aluvisol, the effect of overgrazing in suppressing the development of the root system and, therefore, depleting the soil C (Harris et al., 2007), surpassed the positive fertilizing effect of pasture burning on the accumulation of soil C (Abril et al., 2005).

The total N stocks did not differ between the two pasture systems in the three soil layers (Table 1). The soil under AF showed, in all layers, larger SOC and N stocks, when compared with the pasture environments. These results can be attributed to the high plant residues and shoot production in the forest environment (Backes et al., 2005).

The cumulative C stock up to a 15 cm depth in 1NB was approximately 9.4 Mg ha⁻¹, greater than in 23NB (Table 1), indicating that the soil under frequent burning and low grazing was more efficient at accumulating C than 23NB. The C/N ratio did not differ within the environments except at the 5–10 cm layer, in which 1NB had a higher C/N ratio than 23NB.

In the pasture systems, the C stocks contained in the free light fraction varied from 0.26 to 1.43 Mg ha⁻¹ (Table 2), and the contribution of C_{FLF} to the soil C stock was less than 4% (Figure 1). Regardless of the pasture system, the C_{FLF} stocks decreased with depth (Table 2), as is usually observed (Boeni, 2007). In both layers, the largest values were found in the 23NB system.

The C_{OLF} stocks varied from 4.48 to 8.55 Mg ha⁻¹ in the pasture systems and were approximately 1.9 to 13 times larger than the C_{FLF} stocks (Table 2). These

Table 1. Average±standard deviation of soil organic carbon and N stocks and C/N ratio in bulk soil (C_{soil} and N_{soil}) of the 0–5, 5–10, 10–15, and 0–15 cm layers of a Leptosol under native pasture without burning in the last year (1NB), native pasture without burning in the last 23 years (23NB), and Araucaria forest (AF)⁽¹⁾.

Site	C _{soil} (Mg ha ⁻¹)	N _{soil} (Mg ha ⁻¹)	C/N
		0–5 cm layer	
1NB	33.1±2.9	2.22±0.27	14.9±0.6
23NB	31.2±3.3	2.14±0.54	14.6±2.0
AF	52.0±5.1	3.42±0.27	15.2±0.3
		5–10 cm layer	
1NB	32.8±1.1	2.11±0.16	15.5±0.8
23NB	28.0±1.2	2.13±0.08	13.1±1.0
AF	43.7±4.4	2.93±0.35	14.9±0.6
		10–15 cm layer	
1NB	32.0±3.1	2.04±0.32	15.7±2.3
23NB	26.2±1.1	1.73±0.22	15.1±1.8
AF	41.7±2.0	2.95±0.19	14.1±1.2
		0–15 cm layer	
1NB	97.8±6.9	6.36±0.58	15.4±0.7
23NB	85.4±4.0	6.00±0.66	14.2±1.1
AF	137.4±10.6	9.30±0.38	14.8±0.5

⁽¹⁾Average of three replicates.

results can be associated to the protection of the SOM enclosed in the aggregates, hindering its further microbial decomposition (Conceição et al., 2008). In the 0–5 cm layer, the C_{OLF} under 1NB surpassed the 3.85 Mg ha⁻¹ of C found in 23NB, whereas in the 5–10 cm layer the opposite behavior was observed (Table 2). In the 1NB environment, C_{OLF}/C_{soil} decreased from 26 to 14% with depth (Figure 1), as is normally reported for soils under pasture and under no-tillage (Boeni, 2007). In 23NB, however, this ratio increased from 15 to 19% with depth, indicating that a breakdown of aggregates in the 0–5 cm surface layer occurred due to the higher grazing pressure in this system, promoting a higher decomposition of the OLF.

The C stock contained in the heavy fraction contributed with approximately 72 to 86% of the SOC stock (Figure 1). In the 23NB system, the C_{HF} stock did not vary with depth, but increased from the 0–5 to the 5–10 cm layer in the 1NB system. In the subsurface layer, C_{HF} was greater in the 1NB system than in 23NB (Table 2).

In the 1NB system, a redistribution of the C stocks with depth occurred without a concomitant change of the total SOC (Table 1). The C stocks of both light fractions decreased from the 0–5 to the 5–10 cm layer; however, there was an increase of C_{HF} stocks (Table 2). In the pasture system unaffected by fire for the last 23 years and with a higher grazing intensity, the C_{FLF} and C_{HF} stocks tended to decrease with depth.

Table 2. Average±standard deviation of soil C and N stocks in physical pools and C/N ratio in the 0–5 and 5–10 cm layers of a Leptosol under native pasture without burning in the last year (1NB), native pasture without burning in the last 23 years (23NB), and Araucaria forest (AF)⁽¹⁾.

Site	C (Mg ha ⁻¹) N (Mg ha ⁻¹)		C/N	C (Mg ha ⁻¹) N (Mg ha ⁻¹)		C/N
	(0–5 cm)			(5–10 cm)		
Free light fraction (FLF)						
1NB	0.65±0.06	0.02±0.01	32.5±7.6	0.26±0.07	0.01±0.00	25.7±7.0
23NB	1.43±0.11	0.07±0.01	20.4±4.5	0.90±0.21	0.03±0.01	30.0±6.9
AF	10.81±1.22	0.60±0.14	18.0±5.1	0.90±0.07	0.05±0.01	18.4±3.5
Occluded light fraction (OLF)						
1NB	8.55±1.96	0.49±0.13	17.4±0.6	4.48±0.21	0.25±0.02	18.0±0.9
23NB	4.70±0.59	0.30±0.03	15.5±1.9	5.38±0.19	0.27±0.01	19.7±0.4
AF	13.71±2.12	0.76±0.13	18.1±0.3	10.38±2.00	0.64±0.13	16.3±1.2
Heavy fraction (HF)						
1NB	23.90±0.86	1.70±0.15	14.1±0.8	28.04±0.94	1.84±0.16	15.3±0.9
23NB	25.06±2.98	1.77±0.56	14.7±2.7	21.69±1.54	1.84±0.07	11.8±1.2
AF	27.51±2.90	2.06±0.16	13.4±1.5	32.37±4.98	2.24±0.37	14.5±0.7

⁽¹⁾Average of three replicates.

Among the studied environments, the soil under AF showed the largest stocks of C_{FLF} and of C_{OLF} in the 0–5 cm layer (Table 2), representing 21 and 26% of SOC, respectively (Figure 1). These results can be attributed to the higher input of residues from forest litter, compared with pasture systems (Backes et al., 2005). In the subsurface layer (5–10 cm), the strong reduction of 9.91 Mg ha⁻¹ in the C_{FLF} , in comparison to the above layer (Table 2), was probably due to the lower input of litter at this depth, resulting in a lower proportion of C_{FLF}/C_{soil} (2%) (Figure 1). The C_{OLF} stocks in the 5–10 cm layer remained the highest under AF,

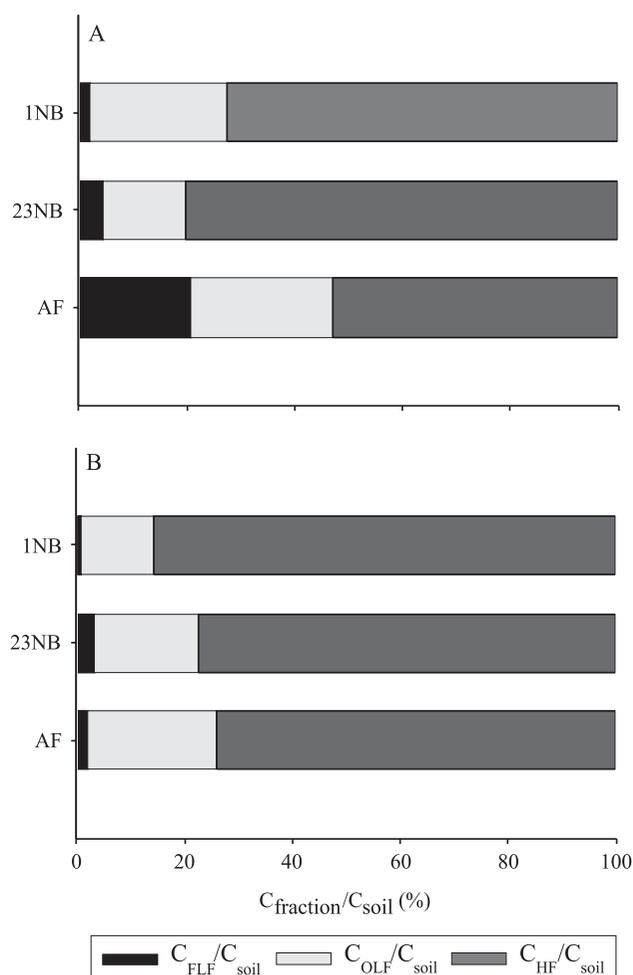


Figure 1. Proportion of soil C stock contained in the free light fraction (FLF) (C_{FLF}/C_{soil}), in the occluded light fraction (OLF) (C_{OLF}/C_{soil}) and in the heavy fraction (HF) (C_{HF}/C_{soil}) C_{OLF} of the 0–5 (A) and 5–10 cm (B) layers of a Leptosol under native pasture without burning in the last year (1NB), native pasture without burning in the last 23 years (23NB), and Araucaria forest (AF).

whereas the C_{HF} was only greater than that found under 23NB. Therefore, the greater capacity of the forest system to accumulate SOM, compared with the pasture systems, occurs mainly due to the SOM accumulation in both light fractions.

The N stocks in the SOM physical pools increased in the order $FLF < OLF < HF$ in all the evaluated systems. In general, no relevant differences between the layers were observed (Table 2). Under AF, the N stocks in both light fractions were larger than those observed in the 23NB system.

The C/N ratio in all systems tended to decrease in the order $FLF > OLF > HF$ in both layers (Table 2), indicating an increase of SOM decomposition in the same order, as expected (Golchin et al., 1994; Tomazi, 2008). The occurrence of microbial products (low C/N ratio) in the HF contributed to its smallest value among the three fractions (Diekow et al., 2005). In general, the C/N ratio in a given fraction did not differ among the pasture systems, and the obtained values were not indicative of black carbon, being within the range observed for light fractions in agricultural soils (Boeni, 2007).

The FTIR spectra of the mineral matrix of the FLF ($FLF_{H_2O_2}$) from the 0–5 cm layer showed the following absorption bands: two peaks at approximately 3,695 and 3,620 cm^{-1} due to the Al–OH inner and outer stretching of kaolinite, bands at approximately 1,084 and 912 cm^{-1} , attributable to the O–Al–OH vibration of kaolinite, and a band at 1,034 cm^{-1} due to the Si–O stretching of quartz (Figure 2). In the $FLF_{H_2O_2}$ spectra from the 1NB and AF environments (Figure 2), a triplet at 3,620, 3,526, and 3,447 cm^{-1} was also identified and was assigned to O–Al–OH vibrations of gibbsite. The same bands, though less intense, were identified in the spectra of the untreated FLF and OLF. These results are evidence of the association of kaolinite and gibbsite (1NB and AF) with the SOM in both light fractions of this soil.

In the spectra of the untreated light fractions (Figure 2), absorption bands resulting from the presence of SOM were detected only at 2,921 and 2,852 cm^{-1} and were assigned to the C–H stretching of aliphatic groups. The hydrofluoric acid treatment removed a great part of the inorganic matrix, and the spectra of the treated fractions (FLF_{HF} , OLF_{HF} , and HF_{HF}) showed the typical pattern of SOM (Figure 2): a band at 3,380 cm^{-1} due to O–H stretching; bands at 2,921 and 2,852 cm^{-1} ; a peak

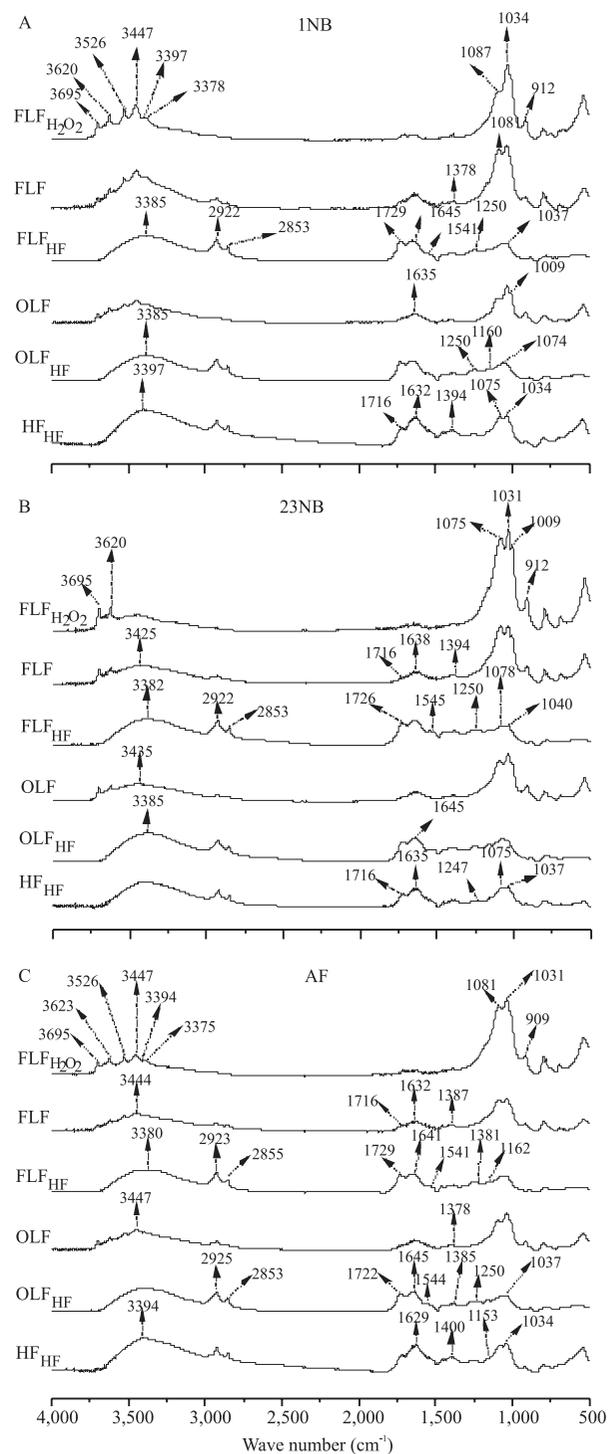


Figure 2. FTIR spectra of the H_2O_2 -treated free light fraction ($FLF_{H_2O_2}$), of the untreated free light fraction and the occluded light fraction (FLF and OLF), and of the hydrofluoric acid-treated OLF and heavy fraction (OLF_{HF} and HF_{HF}) of the 0–5 cm layer of a Leptosol under native pasture without burning in the last year (1NB), native pasture without burning in the last 23 years (23NB), and Araucaria forest (AF).

approximately 1,540 cm⁻¹ attributable to N–H and C=N stretching; a peak at 1,726 cm⁻¹ due to C=O stretching of the COOH group; a band at 1,645 cm⁻¹, resulting from the C=C vibration of the aromatic groups; and a band at approximately 1,235 cm⁻¹ due to C–O stretching of carboxylic acid. The HF_{HF} spectra had a stronger band at approximately 1,075 cm⁻¹ attributable to the C–O stretching of carbohydrate, showing a high proportion of compounds of microbial origin (Wiesmeier et al., 2009). No differences concerning the FTIR spectra of a given fraction were identified among the studied environments.

For T<105°C, the weight loss varied between 1.9 and 7.5% (Table 3), corresponding to the loss of hygroscopic water from the organic matter (Critter & Airoidi, 2006). In general, the largest weight loss occurred in the second (11 to 50%) or third (6.5 to 49%) temperature intervals. The mass loss in the second

interval (105–350°C or 105–425°C) corresponds to the decomposition of higher thermolabile structures, such as alkyl and O–alkyl structures, phenolic and carboxylic groups, and even some alkyl-aromatic groups (Critter & Airoidi, 2006). In the third temperature interval (350–600°C or 425–600°C), less thermolabile structures, such as aromatic structures, decompose (Rosa et al., 2005). The weight loss at 600 to 700°C was very low (≤1.4%) and may be assigned to highly thermostable structures, including condensed aromatic structures. The residual mass for T>700°C was considerably high in the FLF (59 to 80%), corresponding to the inorganic soil matrix adsorbed onto the untreated FLF, as shown by the FTIR spectra (Table 3). For the HF-treated OLF (OLF_{HF}), this residue varied between 13 and 18% and is probably derived from inorganic components of the partially decomposed residues – e.g., Si–O (Silva et al., 2008) –, as indicated by the absorption band at 1,034 cm⁻¹ in the OLF_{HF} FTIR spectra (Figure 2). The HF_{HF} showed a higher proportion of residue mass at T>700°C (18 to 29%), probably due to the high abundance of quartz that is not completely dissolved by the hydrofluoric treatment and due to some neo-precipitated minerals formed during the acid treatment (Dick et al., 2008).

The TGI indicates the proportion between the less and the more thermolabile structures. Considering that thermolability is inversely related to chemical recalcitrance, higher values of the TGI represent a higher chemical recalcitrance of the SOM structure (Benites et al., 2005).

Within the pasture environments, the smallest TGI value of the FLF in the 0–5 cm layer was found in 1NB, indicating a higher proportion of higher thermolabile structures in the SOM of the recently burned environment. The opposite situation was observed with the OLF_{HF} (TGI_{1NB} > TGI_{23NB}), which indicates the predominance of structures with a higher chemical recalcitrance in this fraction. The same behavior occurred with the OLF from the 5–10 cm layer (Table 3), which could be indicative of the occurrence of black carbon. The presence of black carbon in the particulate fraction of agriculture soils was identified in other studies in southern Brazil (Diekow et al., 2005; Brodowski et al., 2006). However, the obtained C/N ratios in the present study did not differ between FLF and OLF (Table 2), and therefore, do not corroborate the occurrence of black carbon in 1NB.

Table 3. Mass loss in four temperature intervals⁽¹⁾, mass residue for T>700°C, and thermogravimetric index (TGI) of soil organic matter physical pools of the 0–5 and 5–10 cm layers of a Leptosol under native pasture without burning in the last year (1NB), native pasture without burning in the last 23 years (23NB), and Araucaria forest (AF).

Site	Mass loss (%)				Mass residue (%)	TGI
	1 st region	2 nd region	3 rd region	4 th region		
Free light fraction (FLF) – 0–5 cm						
1NB	1.93	11.21	6.50	0.09	80.27	0.76 ⁽²⁾
23NB	3.97	19.41	17.11	0.25	59.26	0.99 ⁽²⁾
AF	3.50	20.20	14.12	0.26	61.93	0.73 ⁽²⁾
Occluded light fraction ⁽³⁾ (OLF) – 0–5 cm						
1NB	4.74	32.81	43.56	0.97	17.92	1.33
23NB	5.74	48.65	31.71	0.63	13.27	0.65
AF	5.00	29.88	46.10	1.53	17.49	1.54
Heavy fraction ⁽³⁾ (HF) – 0–5 cm						
1NB	4.61	30.32	39.33	0.04	25.71	1.29
23NB	5.29	31.73	39.50	0.08	23.40	1.24
AF	4.27	32.68	42.84	0.08	20.14	1.31
Occluded light fraction ⁽³⁾ (OLF) – 5–10 cm						
1NB	5.74	33.53	46.83	0.63	13.27	1.40
23NB	4.50	49.94	30.42	0.94	14.20	0.61
AF	- ⁽⁴⁾	-	-	-	-	-
Heavy fraction ⁽³⁾ (HF) – 5–10 cm						
1NB	4.24	28.34	38.63	0.00	28.99	1.36
23NB	7.50	31.16	40.58	1.39	19.37	1.30
AF	5.01	27.66	48.59	0.05	18.69	1.75

⁽¹⁾Temperature intervals for the 1st region: 40–105°C; 2nd region: 105–350°C (1NB and AF) and 105–425°C (23NB); 3rd region: 350–600°C (1NB and AF) and 425–600°C (23NB); 4th region: 600–700°C. ⁽²⁾Thermogravimetric index calculated after subtracting the inorganic matrix contribution. ⁽³⁾Soil physical fractions treated with hydrofluoric acid. ⁽⁴⁾Not analyzed.

The TGI values in the HF were similar in the 1NB and 23NB systems, in both layers (Table 3), indicating that pasture management did not significantly affect the chemical recalcitrance of this fraction. The TGI values for the AF fractions followed the same trend observed for 1NB: in the 0–5 cm layer, it decreased in the order OLF > HF > FLF. With the exception of the FLF, the TGI for a given fraction was always larger under AF, indicating the occurrence of more recalcitrant SOM under forest.

Conclusions

1. Annual pasture burning does not affect soil C stocks up to 15 cm of depth, but promotes a redistribution of soil organic matter (SOM) in physical pools with increasing depth.

2. The frequent burning of vegetation promotes an increase of SOM chemical recalcitrance in the occluded fraction, although black carbon was not identified.

3. The cessation of burning for the last 23 years and a higher grazing intensity promote a decrease of C stocks in the 5–15 cm layer, and a redistribution of the C stock from the occluded fraction to the heavy fraction seems to have occurred in the 0–5 cm layer.

4. The soil under forest shows larger C stocks in the whole soil than in the pasture systems, a behavior that is due to the C accumulation in the light fractions, and SOM chemical recalcitrance is comparatively higher in all SOM physical pools under forest.

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