# EVALUATION OF COPPER RESISTANT BACTERIA FROM VINEYARD SOILS AND MINING WASTE FOR COPPER BIOSORPTION

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Submitted: February 19, 2010; Returned to authors for corrections: March 04, 2010; Approved: June 21, 2010.

#### **ABSTRACT**

Vineyard soils are frequently polluted with high concentrations of copper due application of copper sulfate in order to control fungal diseases. Bioremediation is an efficient process for the treatment of contaminated sites. Efficient copper sorption bacteria can be used for bioremoval of copper from contaminated sites. In this study, a total of 106 copper resistant bacteria were examined for resistance to copper toxicity and biosorption of copper. Eighty isolates (45 from vineyard Mollisol, 35 from Inceptisol) were obtained from EMBRAPA (Empresa Brasileira de Pesquisa Agropecuária) experimental station, Bento Gonçalves, RS, Brazil (29°09'53.92"S and 51°31'39.40"W) and 26 were obtained from copper mining waste from Caçapava do Sul, RS, Brazil (30°29'43.48"S and 53'32'37.87W). Based on resistance to copper toxicity and biosorption, 15 isolates were identified by 16S rRNA gene sequencing. Maximal copper resistance and biosorption at high copper concentration were observed with isolate N2 which removed 80 mg L<sup>-1</sup> in 24 h. Contrarily isolate N11 (*Bacillus pumilus*) displayed the highest specific copper biosorption (121.82 mg/L/OD unit in 24 h). GenBank MEGABLAST analysis revealed that isolate N2 is 99% similar to *Staphylococcus pasteuri*. Results indicate that several of our isolates have potential use for bioremediation treatment of vineyards soils and mining waste contaminated with high copper concentration.

Key words: Copper contamination, vineyard soil; mining waste, copper biosorption; bioremediation.

## INTRODUCTION

Copper is a very important element. Living organisms require copper as an essential micronutrient. Prior to the recognition of the existence of microorganisms on Earth, the Egyptians, Greeks, Romans, and Aztecs used copper compounds for hygiene and for the treatment of diseases (21). Many fungicides, paints, antimicrobial medicines, oral hygiene products, hygienic medical devices, antiseptics and other

products contain copper as an antimicrobial agent (21). However, at high copper concentrations, it is very toxic to most forms of life in addition to microorganisms (5).

Mining activities of modern societies, extensive industrial use of copper and its widespread use as a pesticide in crop production are major sources of copper pollution of soils and water. Toxic heavy metals pose a serious threat to human health, biodiversity and the ecosystem (3). In vineyards, copper pollution negatively impact grape production. Consequently,

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the development of methods to remove toxic heavy metals such as copper from water and soils is currently an area of intensive research (1, 6, 7, 8, 12, 14, 20, 22, 23, 25, 26, 27).

Treatment technologies such as ion exchange adsorption, electrodialyses, precipitation and chemical reduction can be used to remove heavy metals (17). These methods are conversely expensive compared to bioremediation processes. Biological removal of pollutants is attractive in this technology, and it is considered cost-effective and eco-friendly (4, 7, 8, 13, 14, 25). Contaminated environments select resistant microorganisms over pollutants (3). Microorganisms that are resistant to toxic and recalcitrant chemicals can be isolated from polluted sites as well as natural soils and used for bioremediation of environments contaminated with specific chemicals to which they are resistant (6, 7, 15, 27, 28). Biosorption is an important bioremediation process for removal of copper and other toxic heavy metals from the environments. In this study, in search for efficient strains for copper bioremoval, we examined a total of 106 copper resistant bacteria isolated from two copper contaminated vineyard soils and copper mining waste for copper biosorption at high concentration. We employed DNA-based methods to identify promising copper resistant isolates with potential for copper bioremoval from contaminated environments.

## MATERIALS AND METHODS

## Soil sample

Soil samples were collected from three contaminated soils

from South of Brazil. Two of them were collected from copper contaminated vineyards areas of EMBRAPA experimental station, Bento Gonçalves, RS, Brazil (29°09'53.92"S and 51°31'39.40"W). These two soils were classified as Inceptisol and Mollisol. The copper mining waste sample was obtained from the copper mining area of Caçapava do Sul, RS, Brazil (30°29'43.48"S and 53'32'37.87W). Soil samples and copper mining waste were characterized. Table 1 presents the physicochemical parameters analyzed in the Laboratory of Soil Analysis from Federal University of Rio Grande do Sul.

## **Enrichment and Isolation of Copper-Resistant Bacteria**

Enrichment of copper resistant bacteria was in 100 mL of nutrient broth (NB) (5 g of Peptone and 3 g of Beef extract) in 250 mL Erlenmeyer flasks to which 300 mg L<sup>-1</sup> of Cu(II) as copper sulfate (CuSO<sub>4</sub>.5H<sub>2</sub>O) was added and pH was adjusted to 7.0. NB was sterilized by autoclaving at 121°C for 20 min. The soil samples were independently used to inoculate (1%, w/v) sterile medium amended with Cu(II) and incubated for 24 h, with shaking (150 rpm) at 30°C. Subsequently, 1 mL of enrichment culture was used to inoculate 99 mL of sterile medium amended with Cu(II) and incubated for 24 h, with shaking (150 rpm, 30°C). This procedure was repeated two times. Cu(II)-resistant bacterial were thereafter purified by repeated streaking on nutrient agar (NA) plates containing Cu(II) (300 mg L<sup>-1</sup>). The isolates were coded with letter C for Mollisol isolates, letter N for Inceptisol isolates and letter R for waste from copper mining area.

**Table 1.** Chemical and physical properties of vineyard soils contaminated with copper (Inceptisol and Mollisol) and copper mining wastes (Waste).

Tratamento	pН	CEC*	OM**	Clay	Cu	Zn	Mn
	1:1	cmol <sub>c</sub> dm <sup>-3</sup>	g dm <sup>-3</sup>	%		mg dm <sup>-3</sup>	
Inceptisol	6.3	17.2	2.6	19	207	19	55
Mollisol	6.0	13.9	2.5	29	142	18	35
Waste	7.9	-	0.9	2	576	0.8	2
	Ca	Al	Mg	H + Al	S	P	K
		cmolc d	lm <sup>-3</sup>			mg dm <sup>-3</sup>	
Inceptisol	10.9	0.0	3.1	2.8	6.1	28	142
Mollisol	7.8	0.0	2.1	3.5	5.9	27	167
Waste	24.2	0.0	1.7	-	12.3	32	32

<sup>\*</sup> CEC - cation exchange capacity.

<sup>\*\*</sup>OM - organic mater.

# Analysis of isolates for Cu(II)-resistance profile and biosorption

Monoculture isolates were evaluated for Cu(II)-resistance and biosorption as follows: Inoculants were prepared by transferring three loops of each isolate to NB medium amended with 300 mg L<sup>-1</sup> of copper and incubated at 30°C for 24 h with shaking (150 rpm). After, each inoculum was adjusted with sterile saline solution (0.85%) to optical density of 0.85 (OD<sub>600</sub>) and 0.1 mL of each inoculum was added into 20 mL of NB medium containing 300 mg L<sup>-1</sup> of Cu(II) in 50 mL Erlenmeyer and incubated (150 rpm, 24 h, 30°C). Biomass (cell density) was determined by measuring absorbance at OD600 of appropriately diluted cultures. Copper biosorption was determined by measuring copper remaining in the cell-free supernatant, using an atomic absorption spectrophotometer. Briefly, 5 mL of replicate cultures were subjected to centrifugation (10,000 rpm, 10 min). Total copper was analyzed using atomic absorption spectrometer (Perkin-Elmer 2380). Aliquots of culture supernatant (1000 µL aliquots) were diluted 20 times and injected into the atomic absorption spectrometer. Copper biosorption was calculated as the difference in total copper added to the medium and total copper remaining in the medium after different microbial treatments. (CuBiosor = CuTotal added – CuTotal after growth).

## DNA based identification of isolates

Isolates were identified by 16S ribosomal RNA gene sequencing as follows. The isolates were grown by streaking on nutrient agar with incubation at 30°C for 24 h. DNA of each isolate was extracted from colonies forming units pooled from the nutrient agar plate using Promega Wizard Genomic DNA Purification Kit (Promega, Madison, WI) with slight modification. Briefly, cells were re-suspended in 300  $\mu L$  of nucleic acid lyses solution, incubated at 80°C for 15 min and allowed to cool at room temperature. RNase solution (1.5  $\mu L$ ) was added and incubated at 37°C for 60 min. Protein precipitation solution (100  $\mu L$ ) was added and incubated on ice for 5 min. Following centrifugation, the supernatant was transferred to an ice cold tube with 95% ethanol. The

precipitate was recovered by centrifugation. The pellet was washed with 70% ethanol at room temperature and resuspended in sterile nuclease free distilled water. Two primers corresponding E. 27F (5'coli positions AGATTTGATCMTGGCTCAG-3') and 1492R (5'-TACGGYTACCTTGTTACGAC TT-3') were used for PCR amplification of the 16S ribosomal RNA (18). The PCR reaction mixture consisted of 12.5 µL of PCR master mix (Promega, Madison, WI), genomic DNA template (0.5 µL), primer 27F (2.5 µL=12.5 pmol), primer 1492R (2.5 μl=12.5pmol) and made up to 25 μl final volume with nuclease-free water. The 16S rRNA gene was amplified using a 35-cycle PCR (initial denaturation, 95°C for 5 min; subsequent denaturation, 95°C for 0.5 min; annealing temperature, 50°C for 1 min; extension temperature, 72°C for 1 min and final extension, 72°C for 5 min). The PCR amplification products were analyzed by electrophoresis on a 1% agarose gel. Millipore Montage PCR filter units (Millipore, Billerica, MA) were used to remove primers, salts, and unincorporated dNTPs according to the manufacturer's instructions. DNA cycle sequencing was performed using BigDye terminator kit (Applied Biosystems, Foster City, CA) with sequencing primer 519r (5'-GWATTACCGCGGCKGCTG-3') in independent reactions at the Institute of Integrative Genome Biology (IIGB) of UCR, Riverside, CA.

# DNA Sequence Similarity and Phylogenetic Analysis

GenBank BLAST (N) was used for homology searches. Phylogenetic and molecular evolutionary analyses were conducted using MEGA version 4.1 (30). Nucleotide sequence similarity searches were conducted by Genbank BLAST (N). The ribosomal RNA gene sequences were submitted to the GenBank database under accession numbers ranging from FJ577657 to FJ577671.

### RESULTS

# Biosorption of Cu(II) by isolates

Table 2 presents Cu(II) biosorption by 55 bacteria isolated

from Mollisol collected from vineyard soil polluted with copper. Maximum biomass development at high copper concentration (300 mg  $L^{-1}$ ) was observed with isolate C28, C40, C41 and C44. Cu(II) biosorption was maximal in cultures

of isolates C12 (62.21 mg  $L^{-1}$  in 24 h) and C14 (61.77 mg  $L^{-1}$  in 24 h). Isolate C34 displayed the lowest Cu(II) biosorption (6.48 mg  $L^{-1}$  in 24 h) although it grew luxuriantly (1.55 OD<sub>600)</sub> units at high concentration of Cu(II) (300 mg  $L^{-1}$ ).

**Table 2.** Biomass levels and Cu(II) bioremoval in cultures of isolates from vineyard Mollisol incubated in NB medium contaminated with 300 mg L<sup>-1</sup> of Cu(II) and incubated at 30°C for 24 h with orbital shaking.

OD <sub>600</sub> 6±0.0643*	mg L <sup>-1</sup>	mg L <sup>-1</sup> /OD units	
		mg L <sup>-1</sup> /OD units	
	12.53±0.6330	11.73±0.096	
3±0.0464	20.14±1.0338	13.25±0.848	
		20.41±1.890	
		15.76±0.712	
	31.78±2.2971	20.24±1.301	
	35.36±3.1652	32.10±1.762	
		11.12±1.146	
	13.20±0.9496	10.88±0.494	
	36.03±1.5826	26.66±1.208	
	62.21±2.5322	53.71±1.773	
6±0.0250	61.77±2.0675	53.72±2.900	
	29.09±2.9805	36.83±6.043	
		19.82±0.490	
		17.00±1.020	
		33.07±2.614	
		28.83±4.064	
		21.75±0.068	
		21.13±1.781	
		29.09±5.062	
		13.59±0.371	
		51.04±8.672	
		10.59±1.560	
		4.07±0.534	
		6.71±1.746	
		11.01±0.139	
		19.81±0.194	
		9.91±1.303	
		5.79±0.223	
		12.99±1.365	
		10.90±1.057	
		10.12±2.210	
		15.68±2.827	
		22.13±1.710	
		12.20±1.367	
		17.75±3.899	
		9.27±0.596	
		9.67±0.454	
		8.30±1.341	
		11.32±0.902	
		13.85±1.607	
		11.05±3.050	
		16.12±1.155	
		19.79±1.941	
	1±0.0122 4±0.0343 7±0.0328 6±0.0475 5±0.0101 2±0.0250 1±0.0227 9±0.0219 6±0.0250 3±0.0435 5±0.0125 0±0.0195 4±0.0709 3±0.0128 8±0.0554 9±0.0424 1±0.0478 2±0.0080 4±0.0265 5±0.0301 5±0.0242 9±0.0092 7±0.0457 0±0.0168 3±0.0319 2±0.1418 1±0.0269 9±0.0289 5±0.1072 1±0.0347 6±0.0470 0±0.0246 4±0.0962 3±0.0248 6±0.0204 9±0.0205 5±0.1317 5±0.0654 3±0.0335 2±0.0829	1±0.0122    34.91±3.1758      4±0.0343    18.57±0.3165      7±0.0328    31.78±2.2971      6±0.0475    35.36±3.1652      5±0.0101    15.88±1.5826      2±0.0250    13.20±0.9496      1±0.0227    36.03±1.5826      9±0.0219    62.21±2.5322      6±0.0250    61.77±2.0675      3±0.0435    29.09±2.9805      5±0.0125    16.56±0.6330      0±0.0195    13.87±1.2661      4±0.0709    20.14±1.5826      3±0.0128    27.30±3.7983      3±0.0554    34.69±1.5826      9±0.0424    18.79±1.8091      1±0.0478    29.99±5.4273      2±0.0080    25.96±0.6330      4±0.0265    41.40±7.9131      5±0.0301    17.90±2.5322      6±0.0242    6.48±0.9496      9±0.00242    6.48±0.9496      9±0.00457    19.24±1.3675      0±0.0168    31.33±1.1265      3±0.0319    10.29±1.3675      2±0.1418    8.95±1.1265      1±0.0269    24.61±2.3260      9±0.0289    20.59±1.9512      5±0.1697    17.4	

<sup>\*</sup>Values are means ± standard error of the mean

Forty copper resistant bacteria were isolated from Inceptisol collected from copper contaminated vineyard area (Table 3). Growth of the isolates in media amended with 300 mg  $L^{-1}$  was not directly related to copper biosorption. Cell density was highest in cultures of isolate N18 (1.98  $OD_{600}$  units). The highest biosorption of copper was recorded in culture of isolate N2 (80.22 mg  $L^{-1}$  in 24 h) while the lowest Cu(II) biosorption (5.85 mg  $L^{-1}$  in 24 h) was observed for the isolate N20 although cell density was high (1.41  $OD_{600}$  units in 24 h).

Copper biosorption and biomass levels in cultures of 30 bacterial isolates from copper mining waste are presented in Table 4. In general, no direct relationship was observed between amount of biomass in culture and biosorption of Cu(II) by the isolates. Bacterial cell density was highest in cultures of isolates R27 (1.20  $\rm OD_{600}$  units in 24 h), R17 (1.15  $\rm OD_{600}$  units in 24 h) and R8 (1.09  $\rm OD_{600}$  units in 24 h). Maximal Cu(II) biosorption occurred in cultures of isolates R17 (70.47 mg  $\rm L^{-1}$  in 24 h) and R4 (68.34 mg  $\rm L^{-1}$  in 24 h).

**Table 3.** Biomass levels, Cu(II) bioremoval, and specific copper bioremoval in cultures of isolates from vineyard Inceptisol incubated in NB medium contaminated with 300 mg L<sup>-1</sup> of Cu(II) and incubated at 30°C for 24 h with orbital shaking.

Isolates	Biomass	Cu(II) bioremoval	Specific Cu(II) bioremoval
	OD <sub>600</sub>	mg L <sup>-1</sup>	mg L <sup>-1</sup> /OD units
N1	1.46±0.0195*	32.79±1.4768	22.01±0.957
N2	1.45±0.0426	80.22±2.5696	53.40±2.950
N3	1.36±0.0046	32.16±2.3257	23.75±1.776
N4	1.43±0.0231	51.38±1.3427	36.02±1.280
N5	1.13±0.0424	27.15±2.9239	23.65±1.838
N6	1.20±0.0861	30.91±4.1770	27.84±5.232
N7	1.41±0.0288	35.92±3.0121	25.31±1.773
N8	1.38±0.0202	32.16±0.8861	23.14±0.030
N9	1.33±0.0290	22.56±2.4116	16.97±1.805
N10	1.40±0.0150	30.91±0.8980	22.11±0.234
N11	0.67±0.1021	67.25±1.0070	121.83±0.900
N12	$1.44 \pm 0.0358$	29.66±6.4979	21.20±4.386
N13	1.50±0.2895	43.86±4.3875	33.68±2.649
N14	1.61±0.0315	38.01±3.5030	23.53±2.093
N16	1.61±0.0718	44.69±0.5907	27.22±1.822
N17	1.46±0.0396	36.34±4.2869	24.97±2.920
N18	1.98±0.3887	52.21±4.7257	26.10±1.003
N20	1.41±0.0268	5.85±0.5000	4.17±0.112
N22	1.43±0.0478	31.54±2.0675	22.54±2.503
N23	1.40±0.2701	37.80±0.2954	26.96±0.419
N24	1.38±0.2681	37.17±5.3165	26.44±2.927
N25	1.38±0.2647	32.16±7.0886	23.27±5.036
N26	1.33±0.0146	32.16±4.3609	23.96±3.019
N27	1.30±0.0100	19.21±5.5151	14.64±4.152
N28	1.31±0.0070	29.66±6.0242	22.75±4.735
N29	1.36±0.0166	38.01±2.9338	28.10±2.380
N30	$1.41 \pm 0.0031$	16.50±5.0211	11.72±3.553
N32	1.68±0.0959	32.58±5.2559	18.68±2.279
N33	$1.44 \pm 0.0201$	35.30±2.0675	25.07±1.261
N34	1.39±0.0210	7.10±0.5907	4.98±0.366
N35	1.31±0.0177	21.72±5.0356	16.28±3.634
N36	1.29±0.0122	30.91±1.5060	23.92±1.201
N38	1.26±0.0187	35.09±4.6576	27.56±3.245
N39	1.39±0.0181	41.77±1.7390	30.03±1.059
N40	1.18±0.0657	38.84±2.8432	33.48±2.701

<sup>\*</sup>Values are means ± standard error of the mean.

Specific copper removal capacity was generally high in isolates from mining waste. Isolate N11 from vineyard Inceptisol contaminated with copper, however, displayed the highest specific copper bioremoval (121.82 mg/L/OD unit) in 24 h (Table 3). This was followed by isolates R4, R5, R17 and R3 with specific copper bioremoval capacities of 79.70, 67.21,

61.46 and 64.61 mg/L/OD unit in 24 h respectively (Table 4). Statistical evaluation of copper removed by each isolate based on remaining copper in culture supernatant, however, showed that isolate N2 from copper contaminated vineyard Inceptisol was significantly higher than others isolates. Isolates N11, R4 and R17 removed similar levels of copper from the culture.

**Table 4.** Biomass levels and Cu(II) bioremoval in cultures of isolates from copper mining waste incubated in NB medium contaminated with 300 mg L<sup>-1</sup> of Cu(II) and incubated at 30°C for 24 h with orbital shaking.

Isolates	Biomass	Cu(II) bioremoval	Specific Cu(II) bioremoval	
	OD <sub>600</sub>	mg L <sup>-1</sup>	mg L <sup>-1</sup> /OD units	
R1	0.81±0.0009*	35.06±1.1920	43.45±1.528	
R2	0.92±0.0106	34.22±0.6438	37.48±1.236	
R3	0.94±0.0705	56.55±7.7480	36.84±0.900	
R4	$0.80\pm0.0151$	68.35±4.6992	79.70±5.918	
R5	$0.84 \pm 0.0012$	53.60±5.2693	67.21±6.448	
R6	$0.94 \pm 0.0045$	44.33±1.3547	53.06±3.562	
R7	$0.82 \pm 0.0002$	44.75±3.1631	49.54±1.826	
R8	$1.09 \pm 0.0403$	40.75±1.4900	28.98±0.993	
R9	$0.89 \pm 0.0042$	31.27±0.4214	35.46±2.351	
R10	$0.89 \pm 0.0024$	31.69±2.0789	53.35±3.408	
R11	$0.84 \pm 0.0092$	47.28±2.7094	62.70±3.122	
R12	$0.85 \pm 0.0219$	52.34±2.4331	52.10±1.854	
R17	1.15±0.0422	70.46±0.7299	76.49±3.448	
R18	$0.84 \pm 0.0226$	36.74±3.8007	44.77±5.862	
R19	$0.73\pm0.0012$	29.58±0.9733	40.79±1.439	
R20	0.86±0.0415	34.64±1.3547	41.18±3.019	
R21	0.82±0.0019	31.69±1.4800	38.70±1.845	
R22	0.86±0.0099	39.69±1.2875	45.96±1.089	
R23	0.67±0.1426	40.54±1.7032	44.00±1.299	
R24	$0.90 \pm 0.0250$	27.89±0.8773	31.30±0.851	
R25	$0.85 \pm 0.0257$	37.17±2.1212	44.01±3.490	
R26	$0.80\pm0.0073$	30.42±1.7546	38.03±2.596	
R27	1.20±0.0193	36.32±0.8429	30.26±0.718	
R28	0.78±0.0061	30.84±1.9003	39.43±2.634	
R29	0.79±0.0125	27.89±0.9733	36.38±0.935	
R30	$0.85 \pm 0.0047$	23.68±1.1150	27.96±1.447	

<sup>\*</sup>Values are means ± standard error of the mean.

# Identity of selected isolates based on 16S rRNA gene sequence

Fifteen bacterial isolates selected on the basis of copper biosorption and resistances to high copper concentration were identified by 16S rRNA gene sequence analysis. Nucleotide sequences in the range of 471-506 nucleic acid bases (Table 5) were used for Genbank blast analysis and construction of phylogenetic tree. Most of the isolates were identified as

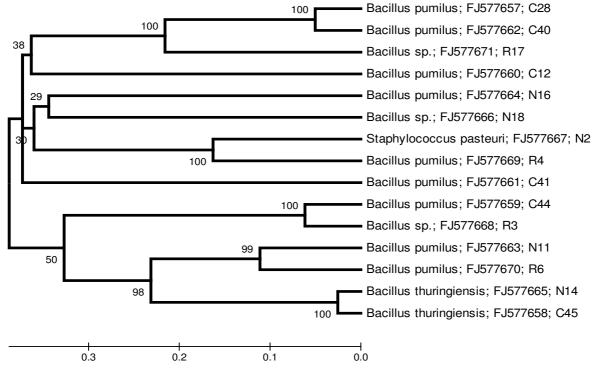
Bacillus species in the phylum Firmicutes. Nine isolates (C28; C44; C12; C40; C41; N11; N16; R4 and R6) were identified as Bacillus pumilus. Two isolates (C45 and N14) were identified as Bacillus thuringiensis and three (N18; R3 and R16) as Bacillus sp. One isolate was identified as Staphylococcus pasteuri (N2). Blast analysis revealed that isolates C44, C41, C40, N11, N2, R4 and R16 were 99% similar to the Genbank match. Isolates C28, C45, C12, N16, N14, N18 and R3 were

98% similar to the Genbank match and 97% similarity was observed for isolate R6. Figure 1 presents the phylogenetic

relationship among selected isolates from three different copper contaminated areas in study.

**Table 5.** DNA-based identification of isolates from different contaminated soils vineyard Mollisol (C), vineyard Inceptisol (N) and copper mining waste (R).

Isolates	Source	16S rDNA Nucleotides	GenBank Submition	GenBank Match	Identity (%)
C28	Mollisol (Vineyard)	494	FJ577657	EU102277.1	B. pumilus (98)
C45	Mollisol (Vineyard)	472	FJ577658	EU037097.1	B. thuringiensis (98)
C44	Mollisol (Vineyard)	476	FJ577659	EF528292.1	<i>B. pumilus</i> (99)
C12	Mollisol (Vineyard)	478	FJ577660	DQ412563.1	<i>B. pumilus</i> (98)
C41	Mollisol (Vineyard)	471	FJ577661	FJ032017.1	<i>B. pumilus</i> (99)
C40	Mollisol (Vineyard)	474	FJ577662	AY792029.1	B. pumilus (99)
N11	Inceptisol (Vineyard)	503	FJ577663	EU102277.1	<i>B. pumilus</i> (99)
N16	Inceptisol (Vineyard)	506	FJ577664	EU855197.1	<i>B. pumilus</i> (98)
N14	Inceptisol (Vineyard)	502	FJ577665	EU037097.1	B. thuringiensis (98)
N18	Inceptisol (Vineyard)	505	FJ577666	EU821778.1	Bacillus sp. (98)
N2	Inceptisol (Vineyard)	494	FJ577667	EU373331.1	S. pasteuri (99)
R3	Mining Waste	502	FJ577668	EU821778.1	Bacillus sp. (98)
R4	Mining Waste	495	FJ577669	EU855197.1	<i>B. pumilus</i> (99)
R6	Mining Waste	495	FJ577670	FM179663.1	<i>B. pumilus</i> (97)
R17	Mining Waste	472	FJ577671	DQ122328.1	Bacillus sp. (99)



**Figure 1.** Phylogenetic tree showing evolutionary distance among selected isolates from three copper contaminated areas based on 16S rRNA gene sequence. The number at each node is the bootstrap from 100 replicates. The scale is the evolutionary distance value.

### **DISCUSSION**

Copper is one of the toxic heavy metals of concern in the environment. Toxicity of heavy metals is largely due to their presence in aqueous systems in ionic forms, which are easily absorbed by living organisms (2, 3). There is increasing interest in the use of microbial biomass for biosorption of heavy metals from the environment. Biosorption of heavy metals involve accumulation of the metals in microbial biomass with subsequent recovery and remediation through bioremediation or chemical technologies. Copper resistant microorganisms with the capacity to adsorb copper on biomass can be used as bioremediation tools to remove copper from contaminated terrestrial and aquatic environments (1).

Several bacteria isolated in this study demonstrated tolerance to extremely high copper concentration. An Agrobacterium tumefaciens, strain CCNWRS33-2 grew in 2.0 mM or 127 mg  $L^{-1}$  of copper in TY liquid medium (5 g tryptone, 3 g yeast extract, and 0.7 g CaCl<sub>2</sub>·2H<sub>2</sub>O per liter), (YMA) (29) but our isolates resisted up to 300 mg L<sup>-1</sup> of Cu(II) in nutrient broth medium, in which they have almost the same composition. The isolate A. tumefaciens CCNWRS33-2 removed 6.35 mg L<sup>-1</sup> of Cu(II) after 36 h (29), and our best isolate (N2) removed 80 mg L<sup>-1</sup> after 24 h, being much more efficient on copper bioremoval. One isolate was identified as Staphylococcus pasteuri. There is little or no published information on copper biosorption by the genus Staphylococcus. However, Staphylococcus warneri was isolated from sediment slurry contaminated with Se (24). Microbial communities are affected by environmental pressures which decrease natural populations and select microbes resistant to contaminants (3). Selenate reducing Bacillus species were abundant in sediment slurries from an evaporation pond heavily polluted with Se and salt (24). Bacillus is an important bacterial genus for bioremediation of heavy metals in different heavy metal contaminated areas (13, 16).

In general, the isolates from the copper mining area showed stronger capacity for copper biosorption. Isolates from copper mining waste area, R4, R17, R3 and R6 removed as much as 70 mg L<sup>-1</sup> from liquid medium in 24 h. A *P. putida* CZ1 (6) and *Pseudomonas* sp. NA (1) removed from liquid medium 20 to 25 mg L<sup>-1</sup> in 24 h. *Staphylococcus pasteuri* N2 isolated in our work from vineyard soil polluted with Cu(II) displayed the highest copper biosorption capacity and removed as high as 80 mg L<sup>-1</sup> of Cu(II) in 24 h. On the contrary, isolate N11 (*B. pumilus*) showed the highest specific rate of copper biosorption calculated relative to cell density. This parity in copper removal capacity of the two isolates indicates that copper biosorption can be directly related to biomass as observed with *S. pasteuri* N2.

High sorptive capacity of prokaryotic (29) and eukaryotic (19) microorganisms are due to components of cell walls that offer an array of functional groups with metal binding capacity. Metals can also be accumulated in the cell cytoplasm (3), but some studies show that higher copper concentrations are linked to the cell wall than the cytoplasm in one prokaryote specie (29). Copper is one of the metals with great potential for bioremoval from contaminated environments through biosorption. In a comparative study (11) on selective binding of different metals to the cell wall of *Pseudomonas* sp., Cu(II) had much more affinity than other heavy metals like Ni(II), Co(II) and Cd(II) when evaluated together.

In summary, we evaluated copper biosorption by several environmental isolates of copper resistant bacteria. DNA-based methods were employed to characterize selected highly copper resistant isolates. Our results showed that several of the isolates have good potential for copper bioremoval from complex environments contaminated with copper.

### **ACKNOWLEDGEMENTS**

Thanks to Brazilian National Research Council (CNPq) for a scholarship to Robson Andreazza and to Auburn University for the opportunity given to Robson to conduct part of his Ph.D. research. An equipment grant to Benedict Okeke to purchase a thermal cycler was received from Auburn University at Montgomery.

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